Accepted Manuscript

Physiological responses and accumulation of heavy metals and arsenic of *Medicago sativa* L. growing on acidic copper mine tailings in arid lands

Fulong Chen, Shuzhi Wang, Shuyong Mu, Iqbal Azimuddin, Daoyong Zhang, Xiangliang Pan, Fahad A. Al-Misned, M.Golam. Mortuza

 PII:
 S0375-6742(15)00113-2

 DOI:
 doi: 10.1016/j.gexplo.2015.05.011

 Reference:
 GEXPLO 5569

To appear in: Journal of Geochemical Exploration

Received date:1 August 2014Revised date:22 April 2015Accepted date:11 May 2015

Please cite this article as: Chen, Fulong, Wang, Shuzhi, Mu, Shuyong, Azimuddin, Iqbal, Zhang, Daoyong, Pan, Xiangliang, Al-Misned, Fahad A., Mortuza, M.Golam., Physiological responses and accumulation of heavy metals and arsenic of *Medicago sativa* L. growing on acidic copper mine tailings in arid lands, *Journal of Geochemical Exploration* (2015), doi: 10.1016/j.gexplo.2015.05.011

This is a PDF file of an unedited manuscript that has been accepted for publication. As a service to our customers we are providing this early version of the manuscript. The manuscript will undergo copyediting, typesetting, and review of the resulting proof before it is published in its final form. Please note that during the production process errors may be discovered which could affect the content, and all legal disclaimers that apply to the journal pertain.



Physiological responses and accumulation of heavy metals and arsenic of

Medicago sativa L. growing on acidic copper mine tailings in arid lands

Fulong Chen^{a,c}, Shuzhi Wang^{a,c}, Shuyong Mu^a, Iqbal Azimuddin^a, Daoyong Zhang^{b*},

Xiangliang Pan^{a*}, Fahad A. Al-Misned^d, M. Golam. Mortuza^{d, e}

^a Laboratory of Environmental Pollution and Bioremediation, Xinjiang Institute of Ecology and Geography, Chinese Academy of Sciences, Urumqi, Xinjiang, 830011, China

^b State Key Laboratory of Environmental Geochemistry, Institute of Geochemistry, Chinese Academy of Sciences, Guiyang, Guizhou, 550002, China.

^c Graduate University of Chinese Academy of Sciences, Beijing, 100049, China

^dDepartment of Zoology, College of Science, King Saud University, Riyadh 11451, Saudi Arabia ^eDepartment of Zoology, Faculty of Life and Earth Science, Rajshahi University, Rajshahi 6205, Bangladesh

*Corresponding author

Prof. Xiangliang Pan Laboratory of Environmental Pollution and Bioremediation, Xinjiang Institute of Ecology & Geography, Chinese Academy of Sciences, Urumqi 830 011, PR China Email: panxl@ms.xjb.ac.cn Tel: +86-991-7885446; Fax: +86-991-7885446

Abstract

Acidic copper mine tailings with high levels of heavy metals and arsenic pose a great risk to ecosystems and human beings. Physiological responses and phytoaccumulation of heavy metals and arsenic in Medicago sativa L. growing in soils with various proportions of acidic copper mine tailings were assessed in this study. Seed germination of *M. sativa* was enhanced by 5-10% tailings but inhibited at higher proportions (30% and 50%) of tailings. Seedling growth, cell membrane and photosynthesis were adversely affected when the plants were grown in soils with high proportions of tailings. The activities of antioxidant enzymes (catalase, superoxide dismutase and guaiocol peroxidase) increased with increasing proportions of tailings, indicating that the antioxidant system was not impaired and can tolerate the toxicity of mine tailings. The plants grew well in soil containing up to 50% of acidic copper mine tailings. Heavy metals and arsenic except Hg were overwhelmingly immobilized in roots. Medicago sativa is a promising plant species for revegetation and phytostabilization of acidic copper mine tailings in arid lands. However, the high mobility of Hg from root to shoot may pose risk to animals.

Keywords: antioxidant enzymes; chlorophyll fluorescence; mine tailings; photosynthesis; phytoremediation; revegetation.

Abbreviation: CAT, Catalase; SOD, Superoxide dismutase; POD, Guaiocol peroxidase

2

1. Introduction

Metalliferous mine tailings contain high contents of heavy metals and metalloids including Pb, Cu, Cd, Hg and As. As one of the major source of heavy metals in the environment, mine waste has caused serious environmental problems in many countries (Bes et al., 2014; Leiva and Morales 2013; Navarro et al. 2008; Pan et al. 2014; Wilson et al., 2005). Recently, China has seen a number of major pollution accidents due to sudden dam break in mine dump ponds or metal leaching from mine refuse heaps in the open. Such pollution disaster has caused pollution of rivers, lakes and soils, and even death of animals and people. The small mine tailing particles and their associated toxic metal(loid)s are prone to wind-borne dispersion and water erosion and have been a great environmental and health risk (Allert et al., 2009; Antunes et al., 2008; Bae et al., 2010; Pyle et al., 2001; Wahl et al., 2012) subsequently requiring cost-effective remediation technologies. The most common way is to simply cap the sites with soil and sand. However, capping with soil and sand is not a permanent solution to this problem because the top layer of the cap is still susceptible to wind and water erosion. Revegetation is considered an inexpensive and ecological friendly strategy for long-term stability of surface layer of mine tailings (Pratas et al. 2005; Tordoff et al., 2000). However, high contents of heavy metals and metalloids, acidity, deficiency of organic nutrients and other harsh environmental factors inhibit plant growth and thus challenge successful application of revegetation. One of the important tasks to overcome these problems is to seek plants that can tolerate the toxicity and grow well in mine tails simply amended with soil. A variety

of metal tolerant plant species have been screened out for revegetation (Abreu et al., 2008; Bech et al. 2012; Lottermoser et al., 2011; Wanat et al. 2013).

The acidic copper mine tailings are most commonly found in northwest China. The tailings are rich in pyrite and heavy metals and metalloids including Pb, Cd, As and Hg. Moreover metal sulfides are prone to oxidation after exposure of the tailings to air and water (Chen and Morris, 1972), and this caused drastic decrease of pH as low as two and leaching out of the associated heavy metals. Owing to the low pH and high levels of heavy metals, few plants can survive in such acidic condition.

In the present study, the potential of alfalfa (*Medicago sativa*) to stabilize acidic copper mine tailings was assessed. There are two reasons for selecting alfalfa as the test plant. One is that it can survive well under various stresses such as drought and low temperature and it can fix nitrogen efficiently, which is important for revegetation in nutrient deficient mine dumps. Seed germination, growth, physiological responses to heavy metals and arsenic of *M. sativa*, and the capacity of this species for phytostabilization of copper mine tailings amended with various proportions of farmland soils were investigated.

2. Materials and methods

2.1 Materials

The copper mine tailings were collected from one copper mine in Xinjiang, China. Farmland soil was collected from Toutunhe Farmland. The soil type is classified as silt loam with 56.5% silt, 25.5% clay and 18% sand. The soil pH, organic matter content and electric conductivity were 7.71, 19.3 g kg⁻¹ and 0.2 mS m⁻¹. Seeds of

Medicago sativa L.cv. Xinmu No.2 was kindly provided by Professor Weijun Li of College of Prataculture and Environment Science, Xinjiang Agricultural University.

2.2 Soil treatment

Farmland soils were mixed with different mass proportions (0%, 5%, 10%, 30% and 50%) of copper mine tailings and contents of heavy metals and pH of the mixed soils were measured, while the mixed soils were filled in pots or Petri dishes for further experiments.

2.3 Seed germination

One hundred alfalfa seeds were sowed in each Petri dish containing soil with various proportions of mine tailings and all tests run in triplicate. The Petri dishes were placed in a plant growth chamber with a day/night humidity of 80%/60%, a day/night temperature of 27/22 °C and a day/night light period of 12h/12h. The soils were watered daily with Milli-Q water and germination percentage was recorded daily. On the 10th day, at least ten seedlings for each treatment were randomly chosen to detect length of hypocotyls and radicle and current weight of seedlings and radicle were also measured.

2.4 Measurement of plant growth

The alfalfa seeds were germinated in Petri dishes at room temperature and were sowed in the soils with various proportions of copper mine tailings. The seedlings were watered with tap water, in which no heavy metals were detected. After three weeks of growth, seedling height was measured. Subsequently, three seedlings were randomly chosen and fresh weight of roots and shoots were measured.

2.5 Chlorophyll fluorescence test

On the third week, OJIP chlorophyll fluorescence of seedling leaves was measured with a pocket fluorometer (FP100, PSI, CZ) (Pan et al., 2011). All leaf samples were dark-adapted for 10 min in advance. Each leaf was tested five times and the mean value was used. F_0 , F_J and F_m from the original measurements were used. $F_{300\mu s}$ are required for calculation of the initial slope (M_o) of the relative variable fluorescence (V) kinetics. The JIP-test parameters were calculated from the OJIP transients (Strasser et al., 2000). The formulae and terms and their illustration used in JIP-test parameters were illustrated in the supplementary Table s1.

2.6 Measurement of chlorophyll content and enzymatic activities

Chlorophyll content of the leaves was measured by spectrophotometry (Lichtenthaler and Wellburn 1983). Briefly, 0.5 g of fresh leaf was extracted in 10 ml of 80% acetone (v/v). The absorbance of the extracts at 663 and 645 for chlorophyll a and b were recorded by a UV-VIS spectrophotometer (UV-2800, Unico, Shanghai, China). Chlorophyll content was calculated according to the equations described by Lichtenthaler and Wellburn (1983).

For analysis of soluble protein content and enzymatic activities, fresh leaves were homogenized in chilled 50 mM potassium phosphate buffer (pH 7.0) containing 2.0% (w/v) insoluble polyvinylpolypyrrolidone, 0.5% (v/v) triton X-100 and 1.0 mM phenylmethylsulfonylfluoride in ice bath then the homogenate was filtered through two-fold muslin cloth and centrifuged at 20,000 ×g for 10 min at 2 °C (Tewari et al., 2006). The supernatant was immediately used for enzyme assays and the soluble

protein content was determined according to the method of Bradford (1976).

Superoxide dismutase (SOD) activity was assayed by monitoring the inhibition of photochemical reduction of nitro blue tetrazolium (NBT) described by Giannopolitis and Ries (1977). The 4 mL of reaction mixture contained 50 mM K-phosphate buffer (pH 7.8), 0.1 mM EDTA, 13 mM methionine, 60 μ M riboflavin, 2.25 mM NBT and an appropriate aliquot of extract. The reaction was measured at 560 nm after 15 min of illumination with light intensity of 75 μ mol m² s⁻¹.

Catalase (CAT) activity was analyzed according to the method described by Aebi (1984). The 3 ml of reaction media contained 30 mM H_2O_2 in a 50 mM phosphate buffer (pH 7.0) and 0.1 ml of enzyme. CAT activity was estimated by monitoring the decrease in absorbance of H_2O_2 at 240 nm using a UV-VIS spectrophotometer (UV-2800, Unico, Shanghai, China).

Peroxidase (POD) activity was measured by the method of Civello et al. (1995). The reaction mixture with a total volume of 3 mL contained 0.02 M Na₂HPO₄-0.08 M NaH₂PO₄ buffer with pH 6.0, 20 mM guaiacol, 4 mM H₂O₂ and 300 μ L of enzymatic extract. The mixture reacted at 30 °C and the enzymatic activity was estimated by measuring the increase of OD (optical density) at 470 nm.

Measurement of proline content in leaves was performed according to the method of Bates et al., (1973). Briefly, 1.0 g of leaf sample was homogenized in 10 mL of 3% sulphosalycylic acid. This extract was used for measurement of proline content spectrophotometrically.

Malondialdehyde (MDA) content in the leaves was measured according to the

method of Rajinder et al (1981). Fresh leaf sample (0.1 g) was homogenized in 5 ml of 0.1% trichloroacetic acid. After 5 min of centrifugation of the homogenate at 10 $000 \times$ g at 4 °C, 0.3 mL of supernatant reacted with 1.2 mL of 0.5% thiobarbituric acid (in TCA 20%) and incubated for 30 min at 95°C. The reaction was terminated in an ice bath for 5 min and the samples were centrifuged at 10 000 × g for 10 min at 25°C. The supernatant absorbance was measured at 532 nm using a UV-VIS spectrophotometer (UV-2800, Unico, Shanghai, China).

2.7 Measurement of heavy metals and arsenic content in soil/mine tailings and plant tissue

For measurement of contents of Cu, Zn, Pb and Cd in soil or/and mine tailings, the samples were ground with an agate mortar and pestle to pass a 100-mesh sieve. One gram of the finely ground sample was digested with HNO₃ and HClO₄ and filtered through 0.45 µm membrane. Contents of Cu, Zn, Pb and Cd in the filtrate were determined by an inductively coupled plasma mass spectroscope (ELANDRC II, PerkinElmer). Another one gram of the finely ground sample was digested with aqua regia and filtered through 0.45 µm membrane. Arsenic and mercury contents in the filtrate were measured by a HG-AFS (AFS-810, Jitan, China). Soil pH was determined at a ratio of soil to deionized water of 1:2.5 using a pH meter.

2.8 Statistical analysis

Each mine tailing treatment was at least triplicated. For biomass, seedling length and other plant growth measurements, at least ten plants were measured. For measurement of enzymatic activities, chlorophyll content and chlorophyll

fluorescence, at least ten plants were measured. All the data used were the mean values. Student's *t*-test was employed for statistical analysis of experimental data. Significance was declared at P < 0.05.

3 Results

3.1 Content of Heavy metals and arsenic in the mine tailings and the soils

Table 1 lists pH and the contents of heavy metals and As of the mine tailings and the farmland soils mixed with various proportions of mine tailings. The copper mine tailings is highly acidic with a pH of 2.8. Cu, Zn, Pb and As contents of mine tailings were several hundreds of mg kg⁻¹. Cadmium and Hg are also far higher than their background values (below 0.5 mg kg⁻¹ for both heavy metals) of most soils (Pérez-Sirvent et al., 2009). The high levels of heavy metals and As in mine tailings imply its high ecological and environmental risk. The farmland soil pH was 7.7 and the content of heavy metals and As of farmland soil were far less than those of mine tailings. The soil pH decreased with increasing proportion of mine tailings mixed. When soil was mixed with 50% mine tailings, soil pH decreased to 6.8. Heavy metal contents in soil increased with increasing proportion of mine tailings.

3.2 Effect of copper mine tailings on seed germination and seedling growth

It can be seen from Fig. 1A that most of the seeds did not germinate during the first four days. The germination percentage for all treatments increased drastically from the fifth day. Most of the seeds germinated at the seventh day. Addition of low proportions (5% and 10%) of copper mine tailings increased the germination percentage by about 10% compared to the control. On 10th d, the maximum seed

germination was found to be 83% for soil containing 10% mine tailings. In comparison with the control, treatment with 30% and 50% mine tailings reduced the germination percentage by about 5% and 10%, respectively. There was no significant difference in germination percentage between different treatments during 1-4 d (p<0.05). Significant difference was found between the control and 50% tailing treatment on 5th and 6th d (p<0.05). There was significant difference between the control and 30% and 50% tailing treatments on 7th d (p<0.05). After 8-9 days of exposure to 30% tailing treatment, germination percentage was significantly different from the control (p<0.1). On 10th d, there was significant difference in germination percentage between all treatments (p<0.1). Generally, seed germination was inhibited by high proportions of mine tailings.

Figure 1B shows length of radicles and hypocotyls of seeds germinated in soil with various proportions of mine tailings. Hypocotyl length for 10% tailing treatment was significantly greater than that for control, 30% and 50% tailings treatments while hypocotyl length for 30% tailing treatment was significantly less than 5% and 10% treatments (p<0.05). Radicle length for 10% tailings treatment was significantly greater than that for 30% tailings treatments (p<0.05) and radicle length for 10% tailings treatment was significantly greater than that for 30% tailings treatments (p<0.05) and radicle length for 50% treatment was significantly lower than for all other treatments.

The biomass of radicles and seedlings growing in soil containing various proportions of mine tailings is shown in Fig. 1C. Biomass of hypocotyl for 10% tailing treatment was significantly higher than that for 30% and 50% treatments, and biomass of hypocotyl for 50% treatment was significantly lower than that for control

and 10% tailing treatment (p<0.05). The radicle biomass for 5% tailing treatment was significantly higher than control, and the radicle biomass for both control and 5% tailing treatment was significantly higher than that for 50% tailings treatment (p<0.05). There was no significant difference between 10%, 30% and 50% tailing treatments (p<0.05).

Root biomass and total length of 31 d old seedlings growing in soil with various proportions of mine tailings are shown in Fig. 1D and 1E. Total seedling length decreased significantly with increasing proportion of tailings from 0 (control) to 30% and there was no significant difference between 30% and 50% tailings treatments (p<0.05). Root biomass for control and 5% and 10% tailing treatments was not significantly different but was significantly higher than that for 30% and 50% tailings treatments (p<0.05). No significant difference in shoot biomass was observed between control and 5% tailing treatment and shoot biomass significantly decreased with increasing proportion of tailings from 5% to 50% (p<0.05), indicating the inhibitory effect of mine tailings on plant growth.

3.3 Effects on biochemical parameters

Soluble protein content of leaf decreased with increasing proportion of mine tailings in farmland soil (Fig. 2A), suggesting that heavy metals and As in the copper mine tailings inhibited synthesis of protein in alfalfa seedlings. 50% mine tailings resulted in a decrease of soluble protein content by 38.4%. The activities of SOD, POD and CAT generally increased with the increase of the proportion of mine tailings (Fig. 2B-2D). Comparing to the control, when the seedlings were treated with 50%

mine tailings, SOD, POD and CAT activities increased by 92.52%, 244.53% and 109.66%, respectively. In comparison with the control, malondialdehyde content in leaf decreased when the seedlings were treated with 5% and 10% mine tailings and increased significantly when the seedlings were grown in soils with higher proportions of mine tailings (Fig.2E-2F). Malondialdehyde content is an indicator of peroxidation of cell membrane lipid. This study indicates that low proportions of mine tailings clearly have damage to the cell membrane lipid system and 30% and 50% mine tailings clearly have damage to the cell membrane lipid system. Proline content decreased when the plants were grown in soils containing 5-30% mine tailings but significantly increased by 30% compared to the control.

3.4 Effects on pigment synthesis and photosynthesis

Relative to the control, the chlorophyll a and b and carotenoid contents increased slightly when the plants were exposed to 5% mine tailings and decreased as the proportion of mine tailings increased (Fig. 2G-H), implying that addition of 5% copper mine tailings stimulates synthesis of pigments while higher proportions of mine tailings inhibits pigment synthesis.

The fluorescence intensity at various steps of the OJIP curves changed with the proportion of mine tailings in soil (Fig. 3), implying that copper mine tailings influenced photosynthesis, especially the electron transport chain in photosystem II (PSII). Table 2 lists some selected JIP-test parameters. It was found that 5% and 10% mine tailings treatments had little effect on Fv/Fm (the maximum quantum yield for primary photochemistry) and other electron transport parameters such as ψo

(probability that a trapped exciton moves an electron into the electron transport chain beyond Q_A) and φEo (quantum yield for electron transport). These parameters were slightly reduced when the plants were grown in soil with 30% or higher proportion of mine tailings, accompanied with increasing DIo/RC (dissipated energy flux per RC) and φDo (thermal dissipation quantum yield), TRo/RC (trapped energy flux per RC), ABS/RC (absorption flux per RC). This finally resulted in a significant decrease of PI_{ABS} (performance index on absorption basis), indicating that the copper mine tailings are toxic to photosynthetic apparatus. The results of chlorophyll fluorescence tests were in consistent with the results of pigments.

3.5 Accumulation of heavy metals in plant tissue

Heavy metals except Hg and As were dominantly accumulated in the roots (Fig. 4). The contents of heavy metals and As in plant tissues generally increased with increasing proportion of copper mine tailings. The contents of Cd, Cu, Zn, Pb and As in roots of plant growing in 50% tailings soil were 7.33, 372.44, 328.13, 5.59 and 9.08 mg kg⁻¹, respectively, which were much higher than those in stems and leaves. When the plants were grown in soil with 50% tailings, the transfer factor from root to leaf of Cd, Cu, Zn, Pb and As was 0.21, 0.084, 0.47, 0.026 and 0.081, respectively. The content of Hg in leaf was far higher than that in root and stem. When the soil was treated with 50% tailings, Hg content in leaf, stem and root was 1.49, 0.27 and 0.15 mg kg⁻¹, respectively. There was significant difference in accumulation of metal(loid)s in root, stem and leaf between different treatments except Pb accumulation in leaf (Fig. 4).

4. Discussion

This study showed that the acidic copper mine tailings contained high levels of Cu, Pb, Zn and As and showed some toxicity to alfalfa due to its high levels of potential toxic elements and low pH (2.8). Although high proportions of mine tailings inhibited seed germination and plant growth (Fig. 1), most plants grew well with high photosynthesis performance in soil mixed with up to 50% mine tailings. This indicates that alfalfa can tolerate the toxicity of multiple heavy metals and As in the mine tailings. Some other plants species can also grow well in various mine tailings or soil contaminated with mine tailings although their physiological indicators were inhibited (Craw et al., 2007; King et al., 2008, Shi et al., 2011). For example, growth of the herbaceous species Sesbania cannabina in Cu tailings an Pb/Zn tailings (Shi et al., 2011), eucalyptus species in arsenical, sulphidic gold mine tailings (King et al., 2008), and Lygeum spartum in acidic mine tailings (pH = 2.9) untreated or treated with CaCO₃ or fertilizers (Conesa et al., 2007) was significantly retarded. However, they were still considered to be suitable for tailings revegetation due to their metal tolerance and high biomass. Similarly, in the present study, alfalfa can also be a potential candidate plant species for revegetation of acidic copper mine tailings.

The acidic copper tailings disturbed the physiological status including protein synthesis, antioxidant system and photosynthetic apparatus of alfalfa. Synthesis of protein was inhibited by tailings. The activities of SOD, POD and CAT increased as the proportion of mine tailings in soil increased (Fig.2B-D). This implies that the reactive oxygen species (ROS) generated due to the toxicity of tailings (Devi and

Prasad 1998; Dietz et al. 1999; Pan et al. 2011), and the increased ROS production can cause inactivation of protein, membrane peroxidation, breakage of DNA strand, and even cell death (Luna et al. 1994). To eliminate the detrimental effects of ROS, a variety of antioxidant enzymes, such as POD, CAT and SOD, are involved in removing ROS (Winston 1990). In the present study, the increase of enzymatic activity means that the antioxidant system was not destructed by the toxicity of up to 50% tailings in soil. The increased MDA content in the presence of high proportions of mine tailings showed that the membrane lipid system was adversely affected to some extent (Fig. 2C-D). However, treatment with low proportions of tailings reduced somewhat proline and MDA content, which at least means that the plant growth is not stressed by low proportions of tailings. Some studies showed similar response patterns of antioxidants system to mine tailings (Jiang et al., 2012; Wu et al., 2011). For example, the activities of SOD and CAT of Paulownia fortunei significantly increased as Cu accumulation increased. However, the seedlings of P. fortunei cultivated in Cu tailings grew poorly due to nutrient deficiency (Jiang et al., 2012). The leaf growth, dry matter accumulation, and photosynthesis of vetiver grass (Vetiveria zizanioides L.) growing in soils treated with high proportions of Pb and Zn tailings were remarkably reduced, whereas accumulation of proline and abscisic acid were enhanced and the activities of superoxide dismutase, peroxidase and catalase were stimulated (Pang et al., 2003).

The data on chlorophyll a and b and carotenoid content and JIP-test analysis showed that pigment synthesis and electron transport flow in PSII were inhibited in

the presence of high proportion of tailings (Fig. 2G-H, Fig. 3), implying that the photosynthetic apparatus is the target site of potential toxic elements. Similarly, Jiang et al. (2012) showed that chlorophyll a and b and carotenoid content of P. fortunei significantly increased for 15.7-78.7 µM Cu treatment and significantly decreased for 157 µM Cu treatment. The JIP test provides valuable information about the fluxes of photons, excitons, electrons and further metabolic events of plants growing in soil with various proportions of mine tailings (Appenroth et al., 2001; Han et al., 2007). Fm value decreased and the shape of J-I-P phase became flat after exposure to high proportions of mine tailings (Fig. 3), implying that PSII reaction centers were partially inactivated and could not be closed and reduction of PQ (non-photochemical phase) was inhibited (Strasser et al. 2000; Zhang et al. 2010). The decrease in Fm associated with the increases of fluorescence intensity at J and I steps suggests that electron transport at the donor side of PSII was inhibited and this leads to accumulation of P680^{*} (Govindjee 1995). The slightly inhibited electron transport parameters and the increased energy dissipation parameters (DIo/RC, \u03c6Do and ABS/RC) for plants treated with high proportion of mine tailings could be attributed to the blocked excitation energy transfer among reaction centers, which finally caused some decrease of PI_{ABS}. A variety of pollutants and mine tailings have similar inhibitory effects on photosynthesis (Lange et al., 2012; Pan et al., 2011; Wang et al., 2012; Zhang et al., 2010; Zhang et al., 2013). These results agree well with the results of seed germination, contents of pigments and the biochemical parameters, suggesting that addition of low proportions of tailings slightly enhanced plant physiological status but

higher proportions of tailings would down-regulate plant physiological status somewhat.

Most of the studied chemical elements, except Hg, were accumulated in roots (Fig. 4), indicating that Medicago sativa L.cv. Xinmu No.2 is a potential phytostabilization plant for Cd, Pb, Cu, Zn and As. Similarly, a number of studies showed that most of heavy metals were accumulated in roots of a variety of plant species that grew in various types of mine tailings or heavy metals polluted soils (Abreu et al. 2012; Bech et al. 2012; Galende et al. 2014; Salazar and Pignata 2014; Trigueros et al. 2012; Wanat et al. 2013). Table 3 listed some examples of accumulation of heavy metals or arsenic in root, stem and leaf of such plants. The plants with translocation factor of metal(loid)s from root to shoot/leaf lower than one may be used for phytostabilization. However, sometimes, even the same plant species may show different metal(loid) accumulation patterns when it grows at different sites. For example, Salazar and Pignata (2014) reported that the translocation factor of Pb from root to shoot of Solanum argentinum Bitter & Lillio growing in Pb contaminated soils at different sites changed from 0.32 to 1.87. This means that it is important to assess the suitability of plants for phytostabilization when they are applied to different sites.

The translocation of Hg from root to shoot might be due to the high levels of chlorides in the tailings. The tailings was recycled using 16 g L^{-1} CuCl₂ to recover Cu, Zn and Ag. In Cl⁻ rich solution, mercury mainly exists in the forms of anionic species such as HgCl₄²⁻ and HgCl₃⁻ owing to the favorable complexation of Hg²⁺ with Cl⁻ (Powell et al. 2005), which may be favorable for transfer of Hg from root to shoot.

However, Hg content in shoots of *M. sativa* is much lower than in other reported plant species that were grown in solution or soils spiked with HgCl₂. Shoots of beard grass has been reported to accumulate Hg up to 13 μ g g⁻¹ growing in 1 mg Hg L⁻¹ solution (De Souza et al. 1999) and <65 μ g g⁻¹ dry weight in shoots from plants growing in soils spiked with 50-1000 μ g g⁻¹ HgCl₂ (Su et al. 2007). Brake fern accumulated 540 μ g g⁻¹ and 1469 μ g g⁻¹ of Hg in shoots after 18 days of growing in soils treated with 500 mg kg⁻¹ and 1000 mg kg⁻¹ of HgCl₂ powder, respectively (Su et al. 2007). This implies that accumulation of Hg in shoots is not only dependent on plant species but also on Hg concentration in the media. In order to assess the potential of phytoextracton of Hg by *M. sativa*, accumulation of Hg in leaves in the presence of a broad concentration range of Hg should be examined and the physiological status of plants should be also monitored in the future study.

In the present study, the good survival of *M. sativa* in soil with high proportions of tailings, overwhelming accumulation of heavy metals and As, except Hg, in root, N_2 -fixing ability and its high tolerance to drought and low temperature stress make *M. sativa* a potential revegetation and phytostabilization plant of mine tailings in arid lands. In engineering application, the inhibitory effects of toxic mine tailings on plants can be alleviated by addition of fertilizers or organic waste such as biosolids (Santibáñez et al., 2008).

5. Conclusions

Seed germination of *M. sativa* was enhanced by low proportions of acidic copper mine tailings but inhibited by higher proportions (>30%) of tailings. Seedling growth,

membrane and photosynthesis were adversely affected at high proportions of tailings in soil. The antioxidation enzymatic activities generally increased with increasing proportions of tailings and this indicates that the antioxidant system was not destructed and can deal with the stress from the toxicity of mine tailings. Regardless of the inhibitory effects of tailings, the plants showed high tolerance to multi-metal(loid) toxicity and grew well in soil mixed with various proportions of highly acidic copper mine tailings. The *M. sativa* can stabilize Cd, Cu and Zn in highly acidic copper mine tailings because of its strong accumulation of heavy metals in root. However, the high mobility of Hg from root to shoot may pose a risk to animals. The differences between the mechanism of translocation of Hg from root to leaf and high accumulation of other heavy metal(loid)s need further study.

Acknowledgements

This work was supported by the National Natural Science Foundation of China (41203088, U1120302 and 21177127). Partial funding for this research was also received from the Visiting Professor Program at King Saud University, Riyadh, Saudi Arabia. We are grateful to the two anonymous reviewers for their valuable comments and suggestions.

References

Abreu, M.M., Santos, E.S., Ferreira, M., Magalhães, M.C.F., 2012. *Cistus salviifolius* a promising species for mine wastes remediation. Journal of Geochemical Exploration 113, 86–93
Aebi, H., 1984. Catalase *in vitro*. Methods in Enzymology 105, 121-126

Allert, A.L., Fairchild, J.F., Schmitt, C.J., Besser, J.M., Brumbaugh, W.G., Olson, S.J., 2009. Effects of

mining-derived metals on riffle-dwelling benthic fishes in Southeast Missouri, USA. Ecotoxicology and Environmental Safety 72, 1642-1651.

- Antunes, S.C., Castro, B.B., Nunes, B., Pereira, R., Gonçalves, F., 2008. In situ bioassay with Eisenia andrei to assess soil toxicity in an abandoned uranium mine. Ecotoxicology and Environmental Safety 71, 620-631
- Appenroth, K.J., Stockel J., Srivastava, A., Strasser, R.J., 2001. Multiple effects of chromate on the photosynthetic apparatus of Spirodela polyrhiza as probed by OJIP chlorophyll a fluorescence measurements. Environmental Pollution 115, 49-64.
- Bae, D.Y., Kumar, H.K., Han, J.H., Kim, J.Y., Kim, K.W., Kwon, Y.H., An, K.G. 2010. Integrative ecological health assessments of an acid mine stream and in situ pilot tests for wastewater treatments. Ecological Engineering 36, 653-663.
- Bates, L.S., Waldren, R.P., Teare, I.D., 1973. Rapid determination of free proline for water-stress studies. Plant and Soil 39, 205-207.
- Bech, J., Duran, P., Roca, N., Poma, W., Sánchez, I., Roca-Pérez, L., Boluda, R., Barceló, J.,
 Poschenrieder, C., 2012. Accumulation of Pb and Zn in *Bidens triplinervia* and *Senecio* sp.
 spontaneous species from mine spoils in Peru and their potential use in phytoremediation. Journal of Geochemical Exploration 123, 109-113
- Bes, C.M., Pardo, T., Bernal, M.P., Clemente, R., 2014. Assessment of the environmental risks associated with two mine tailing soils from the La Unión-Cartagena (Spain) mining district. Journal of Geochemical Exploration. http://dx.doi.org/10.1016/j.gexpl0.2014.05.020
- Bradford, M., 1976. A Rapid and Sensitive Method for the Quantitation of Microgram Quantities of Protein Utilizing the Principle of Protein-Dye Binding. Analytical Biochemistry 72, 248-254.

- Chen, K.Y., Morris, J.C., 1972. Kinetics of oxidation of aqueous sulfide by oxygen. Environmental Science & Technology 6, 529–537.
- Civello, P.M., Martinez, G.A., Chaves, A.R., Anon, M.C., 1995. Peroxidase from strawberry fruit (*Fragaria ananassa* Duch.): Partial purification and determination of some properties. Journal of Agriculture and Food Chemistry 43, 2596-2601.
- Conesa, H.M., Robinson, B.H., Schulin, R., Nowack, B., 2007. Growth of Lygeum spartum in acid mine tailings: response of plants developed from seedlings, rhizomes and at field conditions. Environmental Pollution 145, 700-707.
- Craw, D., Rufaut, C., Haffert, L., Paterson, L., 2007. Plant colonization and arsenic uptake on high arsenic mine wastes, New Zealand. Water, Air and Soil Pollution 179, 351–64.
- Devi, S. R., Prasad, M. N. V., 1998. Copper toxicity in *Ceratophyllum demersum* L. (coontail), a free-floating macrophyte: response of antioxidant enzymes and antioxidants. Plant Science 138, 157–165.
- Dietz, K. J., Baier, M., Kramer, U., 1999. Free radicals and reactive oxygen species as mediators of heavy metal toxicity in plants. In Prasad, M. N. V., Hagemeyer, J. (Eds.), Heavy metal stress in plants-from molecules toecosystems (pp. 73–98). Berlin: Springer-Verlag.
- Galende, M.A., Becerril, J.M., Brarrutia, O., Artetxe, U., Garbisu, C., Hernández, A., 2014. Field assessment of the effectiveness of organic amendments for aided phytostabilization of a Pb–Zn contaminated mine soil. Journal of Geochemical Exploration

http://dx.doi.org/10.1016/j.gexplo.2014.06.006

Giannopolitis, C.N., Ries, S.K., 1977. Superoxide dismutase. I. Occurrence in higher plants. Plant Physiology 59, 309-314

- Govindjee (1995) Sixty-three years since Kautsky: chlorophyll a fluorescence. Australian Journal of Plant Physiology 34, 1073–1079.
- Han, T., Kang, S.H., Park, J.S., Lee, H.K., Brown, M.T., 2008. Physiological responses of *Ulva pertusa* and *U. armoricana* to copper exposure. Aquatic Toxicology 86, 176-84.
- Jiang, Z.F., Huang, S.Z., Han, Y.L., Zhao, J.Z., Fu, J.J., 2012. Physiological response of Cu and Cu mine tailing remediation of *Paulownia fortunei* (Seem) Hemsl. Ecotoxicology 21,759-67.
- King D.J., Doronila A.I., Feenstra C., Baker A.J.M., Woodrow I.E., 2008. Phytostabilisation of arsenical gold mine tailings using four *Eucalyptus* species: Growth, arsenic uptake and availability after five years. Science of Total Environment 406, 35-42.
- Lange, C.A., Kotte, K., Smit, M., van Deventer, P.W., van Rensburg, L., 2012. Effects of different soil ameliorants on karee trees (*Searsia lancea*) growing on mine tailings dump soil-part I: pot trials. International Journal of. Phytoremediation 14, 908-24.
- Leiva, M.A., Morales, S., 2013. Environmental assessment of mercury pollution in urban tailings from gold mining. Ecotoxicology and Environmental Safety 90, 167-173
- Lichtenthaler, H.K., Wellburn, A.R., 1983. Determinations of total carotenoids and chlorophylls a and b of leaf extracts in different solvents. Biochemical Society Transactions 11, 591-592.
- Lottermoser, B.G., Glass, H.J., Page, C.N., 2011. Sustainable natural remediation of abandoned tailings by metal-excluding heather (*Calluna vulgaris*) and gorse (*Ulex europaeus*), Carnon Valley, Cornwall, UK. Ecological Engineering 37, 1249-1253.
- Luna, C. M., Gonzalez, V. S., Trippi, V. S., 1994. Oxidative damage caused by excess copper in oat leaves. Plant Cell Physiology 35, 11–15.

Navarro, M.C., Pérez-Sirvent, C., Martínez-Sánchez, M.J., Vidal, J., Tovar, P.J., Bech, J., 2008.

Abandoned mine sites as a source of contamination by heavy metals: A case study in a semi-arid zone. Journal of Geochemical Exploration 96, 183-193

- Pan, X.L., Zhang, D.Y., Chen, X., Bao, A.M., Li, L.H., 2011. Antimony Accumulation, Growth Performance, Antioxidant Defense System and Photosynthesis of *Zea mays* in Response to Antimony Pollution in Soil. Water, Air and Soil Pollution 251, 517-523.
- Pan, H.J., Zhou, G.H, Cheng, Z.Z., Yang, R., He, L., Zeng, D.M., Sun, B.B., 2014. Advances in geochemical survey of mine tailings project in China. Journal of Geochemical Exploration 139, 193-200.
- Pang, J., Chan, G.S.Y., Zhang, J., Liang, J., Wong, M.H., 2003. Physiological aspects of vetiver grass for rehabilitation in abandoned metalliferous mine wastes. Chemosphere 52, 1559–1570.
- Powell, K.J., Brown, P.L., Byrne, R.H., Gajda, T., Hefter, G., Sjöberg, S., Wanner, H., 2005. Chemical speciation of environmentally significant heavy metals with inorganic ligands. Part 1: The Hg²⁺-Cl⁻, OH⁻, CO₃²⁻, SO₄²⁻, and PO₄³⁻ aqueous systems. Pure Applied Chemistry 77, 739–800.
- Pratas, J., Prasad, M.N.V., Freitas, H., Conde, L., 2005. Plants growing in abandoned mines of Portugal are useful for biogeochemical exploration of arsenic, antimony, tungsten and mine reclamation. Journal of Geochemical Exploration, 85, 99-107.
- Pyle, G.G., Swanson, S.M., Lehmkuhl, D.M., 2001. Toxicity of Uranium Mine-Receiving Waters to Caged Fathead Minnows, *Pimephales promelas*. Ecotoxicology and Environmental Safety 48, 202-214.
- Rajinder, S.D., Dhindsa, P.P., Thorpe, T.A., 1981. Leaf Senescence: Correlated with increased levels of membrane permeability and lipid peroxidation, and decreased levels of superoxide dismutase and catalase. Journal of Experimental Botany 32, 93-101.

- Salazar, M.J., Pignata, M.L., 2014. Lead accumulation in plants grown in polluted soils. Screening of native species for phytoremediation. Journal of Geochemical Exploration 137, 29–36.
- Santibáñez, C., Verdugo, C., Ginocchio, R., 2008. Phytostabilization of copper mine tailings with biosolids: Implications for metal uptake and productivity of *Lolium perenne*. Science of Total Environment 395, 1-10.
- Shi, X., Zhang, X.L., Chen, G.C., Chen, Y.T., Wang, L., Shan, X.Q., 2011. Seedling growth and metal accumulation of selected woody species in copper and lead/zinc mine tailings. Journal of Environmental Science 23, 266–274.
- Strasser, R.J., Srivastava, A,. Tsimilli-Michael, M., 2000. The fluorescence transient as a tool to characterize and screen photosynthetic samples, in: Yunus M., Pathre U., Mohanty P. (Eds.) Probing Photosynthesis: Mechanism, Regulation and Adaptation. Taylor & Francis, London, pp. 443-480.
- De Souza, M.P., Huang, C.P.A., Chee, N., Terry, N., 1999. Rhizosphere Bacteria Enhance the Accumulation of Selenium and Mercury in Wetlands Plants Planta **209**, 259-263.
- Su, Y., Han, F., Shiyab, S., Monts, D.L., 2007. Phytoextraction and Accumulation of Mercury in Selected Plant Species Grown in Soil Contaminated with Different Mercury Compounds. WM'07 Conference, February 25 - March 1, 2007, Tucson, AZ
- Tewari, R.K., Kumar, P., Sharma, P.N., 2006. Magnesium deficiency induced oxidative stress and antioxidant responses in mulberry plants. .Scientia Horticulturae 108, 7-14.
- Tordoff, G.M., Baker, A.J.M., Willis, A.J., 2000. Current approaches to the revegetation and reclamation of metalliferous mine wastes. Chemosphere 41, 219-228.

Trigueros, D., Mingorance, M.D., Oliva, S.R., 2012. Evaluation of the ability of Nerium oleander L. to

remediate Pb-contaminated soils. Journal of Geochemical Exploration 114, 126-133.

- Wahl, J.J., Theron, P.D., Maboeta, M.S., 2012. Soil mesofauna as bioindicators to assess environmental disturbance at a platinum mine in South Africa. Ecotoxicology and Environmental Safety 86, 250-260
- Wang, S.Z., Zhang, D.Y., Pan, X.L., 2012. Effects of arsenic on growth and photosystem II (PSII) activity of *Microcystis aeruginosa*. Ecotoxicology and Environmental Safety 84, 104-11.
- Wanat, N., Austruy, A., Joussein, E., Soubrand, M., Hitmi, A., Gauthier-Moussard, C., Lenain, J., Vernay, P., Munch, J.C., Pichon, M., 2013. Potentials of *Miscanthus giganteus* grown on highly contaminated Technosols. Journal of Geochemical Exploration 126–127, 78–84.
- Wilson, B., Lang, B., Pyatt, F.B., 2005. The dispersion of heavy metals in the vicinity of Britannia Mine, British Columbia, Canada. Ecotoxicology and Environmental Safety 60, 269-276
- Winston, G. W., 1990. Physiological basis for free radical formation in cells: production and defenses In: R. Alscher & J. Cummings (Eds.), Stress Responses in Plants-Adaptation and Acclimation Mechanisms (pp. 57–58). Wiley Liss: New York.
- Wu, Q.H., Wang, S.Z., Thangavel, P., Li, Q.F., Zheng, H., Bai, J., Qiu, R.L., 2011. Phytostabilization Potential of *Jatropha Curcas* L. in Polymetallic Acid Mine Tailings. International Journal of Phytoremediation 13, 8, 788-804.
- Zhang, D.Y., Pan, X.L., Chen, X., Mu, G.J., Li, L.H., Bao, A.M., 2010. Toxic effects of antimony on photosystem II of *Synechocystis* sp. as probed by in vivo chlorophyll fluorescence. Journal of Applied Phycology 22, 479-488.

Zhang, D.Y., Deng, C.N., Pan, X.L., 2013. Excess Ca²⁺ does not alleviate but increases the toxicity of Hg²⁺ to photosystem II in *Synechocystis* sp. (Cyanophyta). Ecotoxicology and Environmental Safety 97: 160–165.

MANNON 4

Table 1 pH and heavy metal and As contents (mg kg⁻¹) of soil mixed with various proportions of copper mine tailings. All the data were presented the mean values \pm S.D. of three measurements.

| | Proportions of copper mine tailings mixed | | | | | | | |
|----|---|-------------|-------------|-------------|--------------|--------------|--|--|
| | 0% | 5% | 10% | 30% | 50% | 100% | | |
| pН | 7.71±0.02 | 7.68±0.10 | 7.60±0.02 | 7.26±0.09 | 6.84±0.05 | 2.80±0.02 | | |
| Cu | 42.60±0.63 | 126.00±3.48 | 196.87±5.32 | 320.17±2.50 | 523.99±14.49 | 979.41±32.22 | | |
| Zn | 106.50±0.92 | 141.19±8.50 | 171.06±6.27 | 323.25±5.26 | 440.81±15.06 | 527.63±60.64 | | |
| Pb | 13.08±0.26 | 21.81±0.59 | 29.69±0.29 | 68.40±1.78 | 258.28±4.82 | 290.12±15.60 | | |
| As | 6.37±0.88 | 21.15±1.79 | 33.50±4.14 | 70.70±6.56 | 97.82±6.84 | 234.43±20.79 | | |
| Cd | 0.40±0.00 | 0.67±0.04 | 0.80±0.04 | 2.01±0.07 | 3.18±0.04 | 4.41±0.08 | | |
| Hg | 0.091±0.012 | 0.144±0.017 | 0.161±0.012 | 0.289±0.033 | 0.263±0.051 | 1.339±0.050 | | |
| | | | | | | | | |

Table 2 JIP-test parameters for seedlings growing in soil containing various proportions of copper mine tailings. Different superscript letters for values in the same row denote significant differences (p<0.05).

The significance decreases in the order of a>b>c.

| | Proportion of copper mine tailings in soil | | | | | | |
|--------------------------------|--|---------------------------|---------------------------|--------------------------|--------------------------|--|--|
| | Control | 5% | 10% | 30% | 50% | | |
| F _v /F _m | 0.822±0.004 ^a | 0.824 ± 0.006^{a} | 0.820±0.006 ^a | 0.805±0.012 ^c | 0.813±0.011 ^b | | |
| ψο | 0.637±0.012 ^a | 0.636±0.021 ^a | 0.613±0.019 ^b | 0.581±0.035 ^c | 0.572±0.022 ^c | | |
| ФЕо | 0.524±0.011 ^a | 0.524 ± 0.019^{a} | 0.502±0.015 ^b | 0.468±0.034 ^c | 0.465±0.021 ^c | | |
| ΦDo | 0.178±0.004 ^c | 0.176±0.006 ^c | 0.180±0.006 ^c | 0.195±0.012 ^a | 0.187±0.011 ^b | | |
| PI _{ABS} | 4.368±0.416 ^a | 4.171±0.585 ^a | 3.611±0.461 ^b | 2.883±0.852 ^c | 2.830±0.524 ^c | | |
| ABS/RC | 1.870±0.073 ^b | 1.985±0.096 ^a | 2.017±0.144 ^a | 2.082±0.166 ^a | 2.092±0.143 ^a | | |
| TRo/RC | 1.538±0.060 ^b | 1.635±0.081 ^a | 1.654±0.123 ^a | 1.675±0.116 ^a | 1.700 ± 0.108^{a} | | |
| ETo/RC | 0.979±0.027 ^{bc} | 1.040 ± 0.049^{a} | 1.012±0.058 ^{ab} | 0.971±0.039 ^c | 0.970±0.047 ^c | | |
| DIo/RC | 0.332±0.015 ^c | 0.350±0.019 ^{bc} | 0.363 ± 0.024^{b} | $0.407{\pm}0.054^{a}$ | 0.393±0.043ª | | |
| | \mathbf{O} | | | | | | |

Table 3 List of heavy metals or arsenic contents in root, stem and leaf of plants accumulate metal(loid)s mainly in root in the literature (mg kg⁻¹ dry weight).

| Plant species | Soil/tailings | Metal(loid)s | Root | Shoot/Sten | n Leaf | References |
|------------------------|--------------------------------------|--------------|-----------------|------------|----------|-----------------------|
| Nero oleander | soil with 270 mg kg ⁻¹ Pb | Pb | 13 (inner part) | | 2 | Trigueros et al. 2012 |
| | | | 26 (root bark) | S- | | |
| Miscanthus giganteus | gold mine soils | As | 602.4-1284.5 | 3.6-5.4 | 4.3-17.1 | Wanat et al. 2013 |
| | | Sb | 17.9-26.6 | 0.8-1.0 | 0.2-1.1 | |
| | | Pb | 38.2-326.7 | 0.6-29.5 | 1.1-43.1 | |
| Bidens triplinervia L. | polymetallic mine soil | Pb | 5187 | 655 | | Bech et al. 2012 |
| | (mainly Ag, Pb and Cu) | Zn | 9900 | 1614 | | |
| Sorghum halepense (L. |) Pb contaminated soil | Pb | 1406.8 | 3.1 | 5.2 | Salazar and Pignata |
| Pers. | 0 | | | | | 2014 |
| Medicago sativa L. | soil with 50% acidic Cu | Cd | 7.3 | 2.2 | 1.5 | This study |
| | mine tailings | Cu | 372.4 | 23.6 | 31.5 | |
| 0 | | Zn | 328.1 | 157.8 | 154.4 | |
| | | Pb | 5.6 | 0.6 | 0.1 | |
| | | As | 9.1 | 1.37 | 0.7 | |
| | | Hg | 0.15 | 0.27 | 1.48 | |

Figure legends

Fig. 1 (A) Alfalfa seed germination (%) in soil treated with various proportions of copper mine tailings as a function of time; (B) Length of hypocotyl and radicle of alfalfa seeds germinated in soils with various proportions of mine tailings; (C) Biomass of the seedlings and radicles grown in soils with various proportions of mine tailings; (D) Fresh biomass of the seedlings treated with various proportions of mine tailings. (E) Total length of the seedlings treated with various proportions of mine tailings in soil. * near the bars in Fig. 1A denotes significant difference at p=0.05 or p=0.1 (Duncan's New Multiple Range Test). Different letters above the bars in Fig. 1B-1E denote significant difference at p=0.05 (Duncan's New Multiple Range Test).

Fig. 2 Contents of soluble protein, proline and MDA, enzymatic activities and pigments of the seedlings growing on soils with various proportions of copper mine tailings. (A) the soluble protein content; (B) SOD; (C) POD; (D) CAT; (E) proline content; (F) MDA content; (G) chlorophyll content; and (H) carotenoid content. Different letters above the bars denote significant difference at p=0.05 (Duncan's New Multiple Range Test).

Fig. 3 OJIP transients of seedlings growing on soils with various proportions of copper mine tailings. Fig. 4 Contents of Cd (A), Cu (B), Zn (C), Pb (D), As (E) and Hg (F) in roots, shoots and leaves of the alfalfa seedlings growing on soils with various proportions of copper mine tailings. The symbol asterisks (*)above the bars denote significant difference at p=0.05 (Duncan's New Multiple Range Test).

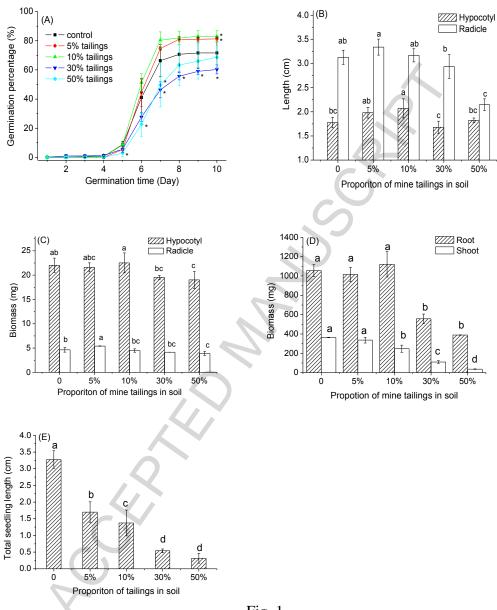
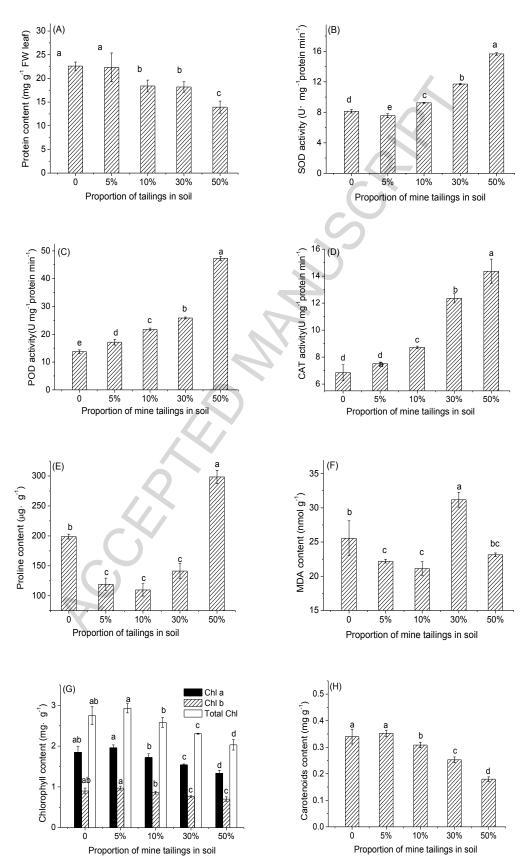
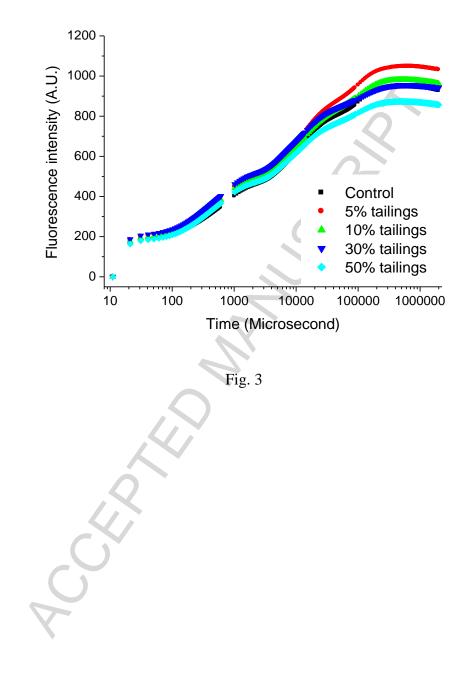
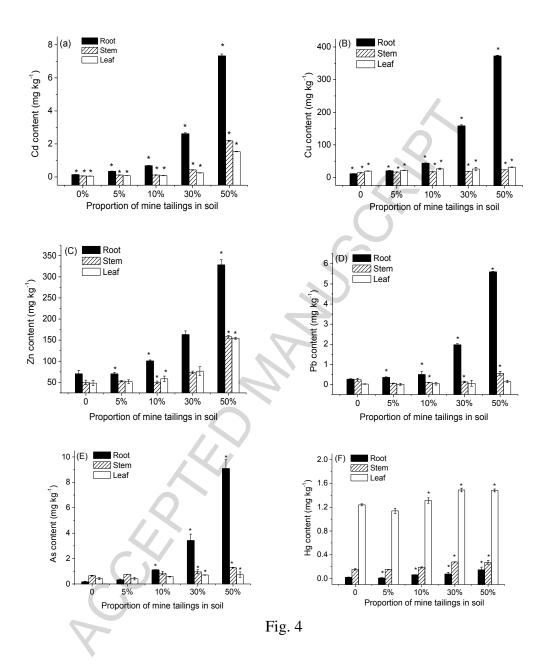


Fig. 1









EPTED MANU

Highlights

- CAT, SOD and POD activities increased with increasing levels of tailings in soil" •
- Arsenic and heavy metals, except Hg, were overwhelmingly immobilized in root
- M. sativa is a potential phytostabilization species of acidic copper mine tailings

<text>